

1 **Variation of the photosynthetic electron transfer rate and**
2 **electron requirement for daily net carbon fixation in Ariake**
3 **Bay, Japan**

4 **Y. Zhu ^a, J. Ishizaka ^b, S.C. Tripathy ^c, S. Wang ^d, Y. Mino ^b, T. Matsuno ^e, D.J. Suggett ^f**

5 ^a Graduate School of Environmental Studies, Nagoya University, Nagoya 464-8601, Japan

6 ^b Institute for Space-Earth Environmental Research, Nagoya University, Nagoya 464-8601, Japan

7 ^c National Centre for Antarctic and Ocean Research, Earth System Science Organization, Ministry of Earth
8 Sciences, Vasco-Da-Gama, Goa 403804, India

9 ^d School of Marine Sciences, Nanjing University of Information Science and Technology, Nanjing, 210044,
10 Jiangsu, China

11 ^e Research Institute for Applied Mechanics, Kyushu University, Fukuoka, Japan

12 ^f Plant Functional Biology and Climate Change Cluster, University of Technology Sydney. P.O.Box 123, Broadway,
13 NSW 2007, Australia

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16

17 **Abstract**

18 Fast Repetition Rate fluorometry (FRRf) provides a potential means to examine marine primary
19 productivity; however, FRRf-based productivity estimations require knowledge of the electron
20 requirement (K) for Carbon (C) uptake (K_C) to scale an electron transfer rate (ETR) to CO_2 uptake
21 rate. Most previous studies have derived K_C from parallel measurements of ETR and CO_2 uptake
22 over relatively short incubations, with few from longer-term daily-integrated periods. Here we
23 determined K_C by comparing depth-specific, daily ETRs and CO_2 -uptake rates obtained from 24-h
24 on-deck incubation experiments undertaken on seven cruises in Ariake Bay, Japan from
25 2008-2010. The purpose of this study was to determine the extent of variability of K_C and to what
26 extent this variability could be reconciled with the prevailing environmental conditions, and
27 ultimately develop a method for determining net primary productivity (NPP) based on FRRf
28 measurements. Both daily ETR and K_C of upper layer varied considerably, from 0.5 to 115.7
29 $mmol\ e^- mgChl-a^{-1}\ d^{-1}$ and 4.1 to 26.6 $mol\ e^- (mol\ C)^{-1}$, respectively, from throughout the entire
30 dataset. Multivariate analysis revealed a strong correlation between daily photosynthetically active
31 radiation (PAR) and K_C ($r^2 = 0.94$). A simple PAR-dependent relationship derived from the data set
32 was used for generating K_C and this relationship was validated by comparing FRRf predicted NPP
33 with ^{13}C uptake measured in 2007. These new observations demonstrate the potential application
34 of FRRf for estimating regional NPP from ETR.

35

36 **Keywords:** FRR fluorometry; primary productivity; ETR; quantum requirement for carbon
37 fixation; ^{13}C -uptake

38

39 **1. Introduction**

40 Fast Repetition Rate fluorometry (FRRf) non-destructively characterizes
41 photosynthetic processes, including photosystem II (PSII) photochemistry and
42 photosynthetic electron transport (Kolber et al., 1998). The technique has been widely
43 considered a key milestone in aquatic research for global efforts to understand
44 environmental regulation of primary productivity since the data can be acquired over
45 unprecedented time and space scales, and with a resolution far beyond that from
46 conventional incubation methods (Suggett et al., 2009a). While FRRf has been
47 principally applied to phytoplankton studies examining effects of physiological stress,
48 such as nutrient limitation (e.g. Behrenfeld and Kolber, 1999; Moore et al., 2006;
49 Moore et al., 2008), it has also been utilized for characterisation light absorption
50 (Suggett et al., 2004; Silsbe et al., 2015) and for interpreting phytoplankton
51 photo-physiological processes in the context of phytoplankton community structure
52 (Suggett et al., 2009b). The most attractive application is its potential for describing
53 primary productivity (e.g. Suggett et al., 2001; Fujiki et al., 2008; Cheah et al., 2011).
54 FRRf thus offers the potential for ground-truthing of remote sensing-based models for
55 estimating primary productivity that are presently poorly resolved over time and space
56 by conventional *in-situ* or simulated *in-situ* incubation CO₂ uptake experiments (Saba
57 et al., 2011; Tripathy et al., 2012).

58 Determination of photosynthetic rate via FRRf differs from conventional
59 CO₂-uptake (or O₂ evolution) measurements since it is derived from simultaneous
60 measurements of light harvesting and utilization by photosystem II (PSII) (Kolber and
61 Falkowski, 1993; Suggett et al., 2015) to yield PSII electron transfer rates (ETR,
62 Oxborough et al., 2012). However, knowledge of the “quantum requirement for
63 carbon fixation” ($\Phi_{e,C}$; Lawrenz et al., 2013, but recently termed K_C ; Hancke et al.,

64 2015) is essential to convert the ETR to a biogeochemically meaningful measure of
65 primary productivity (Suggett et al. 2009a; Lawrenz et al., 2013; Hancke et al., 2015).

66 K_C implicitly accounts for various factors that decouple the ETR from CO_2
67 uptake, notably cellular processes other than CO_2 assimilation that consume energy
68 and reductant (e.g. Suggett et al., 2009a). However, since K_C is determined from
69 parallel measurements of ETR and CO_2 uptake, it also accounts for methodological
70 limitations inherent to estimating the ETR (Schuback et al., 2015) and CO_2 uptake
71 (Kromkamp et al., 2008; Lawrenz et al., 2013). Even so, such ‘conversion factors’
72 appear generally robust since they can be modeled as a function of key environmental
73 drivers across local to regional scales (Lawrenz et al., 2013; Schuback et al., 2015).

74 Lawrenz et al. (2013) synthesized available global FRRf-based K_C estimates to
75 demonstrate that K_C could be frequently predicted as a function of key environmental
76 factors controlling primary productivity, i.e. light, temperature and inorganic nutrient
77 availability, albeit with large spatial variability. Whilst the reason underpinning this
78 variability still remains unclear, it likely reflects localized differences in physiological
79 conditions regulated by specific phytoplankton communities and their environmental
80 conditions. Indeed recent efforts of sampling across broad iron-limited regions of the
81 northeast subarctic Pacific indicate that K_C is highly predictable and the estimation
82 can in fact be achieved from other FRRf-based physiological proxies, such as
83 non-photochemical quenching (Schuback et al., 2015).

84 Despite the increasing use of FRRf within the coastal waters of Japan, there have
85 been no focused studies aimed at understanding the variability of K_C and its regulation
86 by environmental factors. Here we report new data describing the variability of K_C
87 from a semi-enclosed bay, Ariake Bay, Japan. This study location (Fig. 1) offered us
88 the opportunity for repeated sampling under a range of environmental conditions.

89 Ariake Bay is nutrient rich and considered a highly productive ecosystem (Ishizaka et
90 al., 2006). Previous studies in this bay by Tripathy et al. (2010, 2012) provided strong
91 indications that the FRRf might be a powerful tool for estimating primary productivity
92 in this area, but was limited by the need to assume K_C . Here we propose a new
93 FRRf-based approach for examining K_C for this region by comparing daily-integrated
94 ^{13}C -uptake rates (and hence net primary productivity) with corresponding
95 measurements of daily-integrated ETRs. The main objectives of this study are three
96 fold: first to determine the extent of variability of K_C ; second, to evaluate the
97 environmental factors responsible for this variability and third, to establish an FRRf
98 based approach for *in situ* net primary productivity (NPP) estimation.

99

100 **2. Materials and Methods**

101 **2.1. Sample collection, processing and analyses**

102 The sampling protocols utilized in our study are similar to those reported in
103 Tripathy et al. (2010; 2012) and Shibata et al. (2010). In brief, the core observations
104 were from 7 cruises in Ariake Bay (Fig. 1) during fall/spring seasons of 2008, 2009
105 and 2010 (Table 1). An additional sampling campaign from December 2007 was used
106 for later validation of our FRRf-NPP prediction model. For all cruises, seawater
107 samples were collected at a fixed station “T1” (130.31° E, 32.91° N, Fig. 1) from 6
108 discrete depths (corresponding to ca. 1%, 5%, 10%, 25%, 50% and 100% of
109 Photosynthetically Active Radiation, PAR, 400-700 nm, at the sea surface) using a
110 rosette sampler equipped with twelve 5 L Niskin bottles (General Oceanics) and a
111 conductivity-temperature-depth profiler (CTD, 911+, SeaBird Electronics). Data from
112 the CTD was utilized for establishing hydrographic conditions at St. T1 at the time of

113 sampling. For 24 h on-deck ^{13}C uptake incubations, and for Chl-a, nutrient and
114 phytoplankton spectral light absorption measurements, water samples were taken
115 from pre-dawn CTD casts. All seawater samples were first carefully drained from the
116 Niskin bottles into appropriate incubation and/or sampling bottles and further
117 processed as described below.

118 Chl-a concentrations were determined from 100 ml seawater samples filtered
119 onto 25 mm glass fiber filters (Whatman GF/F) under low vacuum (<0.02 MPa).
120 Filters were then extracted in N, N-dimethylformamide for 24 hours in darkness
121 (Suzuki and Ishimaru, 1990) and Chl-a quantified using a pre-calibrated fluorometer
122 (10-AU, Turner Design). Samples for nitrate + nitrite, phosphate and silicate analyses
123 were stored at -20 °C until later analysis using an automated nutrient analyzer
124 (AACS-IV, BLTEC).

125 Phytoplankton absorption coefficients, $a_{\text{ph}}(\lambda)$ (m^{-1}), were determined using the
126 quantitative filter technique of Kishino et al. (1985) as adapted by Wang et al. (2014).
127 250 ml seawater samples were carefully filtered through Whatman GF/F (25 mm) to
128 ensure even distribution of the particulate matter on the filters. The optical density
129 (OD) of all particulate components were measured in a dual beam multi-purpose
130 spectrophotometer (MPS-2400, Shimadzu Inc.) over wavelengths from 350 to 750 nm
131 at 1-nm intervals, before and after the extraction of phytoplankton pigments by
132 methanol. The absorption spectra of particulate material obtained after pigment
133 extraction is that of non-phytoplankton particles. The spectral absorption coefficients
134 of total particulate material ($a_{\text{p}}(\lambda)$) and non-phytoplankton particles ($a_{\text{np}}(\lambda)$),
135 respectively) were determined using the equation $2.303 \cdot \text{OD}(\lambda) \cdot S/V$, where 2.303 is
136 the factor used to convert common to natural logarithm, S is the filter clearance area
137 (m^2), and V is the volume of sample water filtered (m^3). Phytoplankton specific

138 absorption spectra ($a_{ph}(\lambda)$) were then obtained by subtracting the non-phytoplankton
139 absorption from total absorption as,

$$140 \quad a_{ph}(\lambda) = a_p(\lambda) - a_{nph}(\lambda) \quad (1)$$

141 Chl-a specific absorption coefficient, $a_{ph}^*(\lambda)$ ($m^2 \text{ mg Chl-a}^{-1}$), was then calculated by
142 standardizing $a_{ph}(\lambda)$ by Chl-a concentration.

143 **2.2. ^{13}C uptake based net primary productivity estimates**

144 ^{13}C uptake experiments were carried out via 24 h on-deck simulated-in-situ (SIS)
145 incubations with enrichment of ^{13}C stable isotope (min 98 atom%; $\text{NaH}^{13}\text{CO}_3$,
146 ISOTECH), where the final ^{13}C atom % of total dissolved inorganic carbon was ca. 10%
147 of that in the ambient water (Hama et al., 1983). ^{13}C labelled sodium bicarbonate was
148 added to each of quadruplicate 1 L bottles containing seawater from each light depth
149 sampled; three of the replicates were then immediately transferred to incubators in
150 which light was attenuated with blue plastic filters (The General Environmental
151 Technos), and the fourth replicate filtered immediately to serve as the time zero, and
152 adsorption correction of the ^{13}C label on to the filters. The use of blue filters in the
153 incubators simulated *in-situ* irradiance spectra at depths consistent with sample
154 collection. Incubators were set up with a continuous flow-through system of surface
155 seawater to maintain ambient temperatures. We acknowledge that differences between
156 incubation and *in situ* temperatures for deeper samples may raise errors for carbon
157 uptake rates; however, these errors should not be significant in this study since
158 temperature differences between surface and Z_{eu} were less than $0.7 \text{ }^\circ\text{C}$ across all
159 cruises (Davison, 1991).

160 All samples were filtered through 25mm pre-combusted GF/F filters ($450 \text{ }^\circ\text{C}$, 4
161 hours) and stored at $-20 \text{ }^\circ\text{C}$ until further analysis. The filter samples were vacuum

162 dried after exposure to fumes of HCl to remove excess inorganic particulate carbon.
163 The concentration of particulate organic carbon (POC) and the isotopic ratio of ^{13}C
164 and ^{12}C (^{13}C atomic %) on the filters were then determined by isotope ratio mass
165 spectrometer (Delta^{PLUS}, Thermo Fisher Scientific) equipped with an elemental
166 analyzer (EA 1110, CE Instruments). Carbon fixation rates (PP) were calculated
167 according to Hama et al. (1983). Chl-a concentrations before and after the incubation
168 were measured to examine for any drift during the incubations; Chl-a decreased
169 significantly (from ca. 4.43 to 2.28 mg m⁻³) between the start and end of the
170 incubation period from the cruise on the 14 May, 2010, and so the primary production
171 data from this cruise were excluded from all further analysis. Chl-a specific primary
172 productivity (P^B) was PP divided by Chl-a concentration, and the water column
173 integrated PP (IPP) was defined as $\int_0^{Z_{eu}} \text{PP}(z) dz$.

174 **2.3 *in-situ* PAR and irradiance spectrum measurements**

175 Underwater PAR was quantified using a scalar sensor (QSP-200L and QSP-2200,
176 Biospherical Inc.) attached to both the CTD and FRRf frames, and these datasets were
177 both used to subsequently determine the diffuse attenuation coefficient for PAR, $K_d(z)$
178 (m^{-1}). Incident PAR at the sea surface ($\text{PAR}(0^+)$) was measured throughout the
179 incubation period using a quantum scalar irradiance sensor (QSL-2100, Biospherical
180 Inc.) mounted on the incubators used for ^{13}C uptake experiments. The *in situ*
181 underwater irradiance field $E_d^-(\lambda, z)$ was measured by an underwater spectral
182 radiometer (PRR-800, Biospherical Instruments) at 13 wavelengths ($\lambda = 380, 412, 443,$
183 $465, 490, 510, 532, 555, 565, 589, 625, 665$ and 683 nm) every 2 hours from
184 8:00-16:00 (local time).

185

186 2.4. FRRf parameters, electron transport rate (ETR) and K_C

187 Daily time series of active Chl-a fluorescence profiles were measured every 2
188 hours from dawn until dusk (Table 1) from the near surface to depths of ca. 20 m in
189 the water column using a Diving Flash FRRf (Kimoto Electric). The profiling speed
190 was set to $< 0.2 \text{ m s}^{-1}$ to ensure acquisition of fine scale vertically resolved active
191 fluorescence data (Mino et al., 2014). The Diving Flash is designed to measure
192 fluorescence simultaneously in a “dark” chamber, which is fully shaded, and in a
193 “light” chamber equipped with a dichroic cyan filter to reduce the probable influence
194 of red light effect (Raateoja et al., 2004; Tripathy et al. 2010). The FRRf uses a
195 single-turnover protocol and was programmed to deliver sequences of 60 saturation
196 blue flashlets (wavelength of 470 nm with a 25 nm half bandwidth, each 2 μs in
197 duration at 4 μs intervals), followed by 20 relaxation blue flashlets (each 2 μs in
198 duration at 75 μs intervals) (as per Fujiki et al., 2008). Fluorescence parameters
199 derived from both dark (F_o , F_m , σ_{PSII} and F_v/F_m) and light (F' , F_m' , σ_{PSII}' and
200 F_q'/F_m') chambers (refer to Table 2) were calculated using the software (FRRCalc2,
201 Kimoto). Calculation was based on the KPF model fits to FRRf induction curves to
202 resolve physiological parameters (Kolber et al., 1998). Inter-calibration of dark and
203 light chambers was undertaken based on the F_m correction from pre-dawn casts.

204 Measurements of the effective absorption cross section (σ_{PSII}) were weighted to
205 the narrow blue excitation waveband of the FRRf light source (470 nm) and therefore
206 were adjusted to the spectra of the *in situ* irradiance following Suggett et al. (2006b)
207 as:

$$208 \quad \sigma_{PSII}^{abs} = \left(\frac{\bar{a}^{chl}(in\ situ)}{\bar{a}^{chl}(FRRf)} \right) \times \sigma_{PSII}, \quad (2)$$

209

210 where $\bar{a}^{chl}(FRRf)$ and $\bar{a}^{chl}(in\ situ)$ represent the absorption coefficients from

211 Eq.(1) weighted to the incident spectra of the FRRf's excitation (Suggett et al., 2004)
 212 and *in situ* irradiance, respectively. $\bar{a}^{chl}(FRRf)$ and $\bar{a}^{chl}(in\ situ)$ were calculated
 213 from $\bar{a}^{chl} = \sum([a^{chl}(\lambda) \times E(\lambda)] / \sum[E(\lambda)])$, where E refers to the irradiance
 214 spectrum of either the FRRf excitation LED or *in situ* light field (see Suggett et al.,
 215 2001). The spectral range λ of $\bar{a}^{chl}(FRRf)$ and $\bar{a}^{chl}(in\ situ)$, 400 – 700 nm, was
 216 assessed at 1 nm intervals and therefore 1nm resolved *in situ* light fields were
 217 generated by interpolating between the discrete wavebands of the PRR800 (Suggett et
 218 al., 2001).

219 Absolute electron transfer rate (ETR, mmol e⁻ (mg Chl-a)⁻¹ h⁻¹) were then
 220 determined following Suggett et al. (2009a) as follows:

$$221 \quad \text{ETR} = \text{PAR} \times \sigma_{PSII}^{abs} \times n_{PSII} \times q_p \times 0.0243, \quad (3)$$

222
 223 where PAR is in units of $\mu\text{mol quanta m}^{-2} \text{ s}^{-1}$, σ_{PSII}^{abs} is the spectrally corrected
 224 effective absorption cross section of PSII from the dark chamber ($\text{\AA}^2 \text{ quanta}^{-1}$),
 225 and n_{PSII} is the ratio of functional PSII reaction centres to Chl-a (mol RCII (mol
 226 Chl-a)⁻¹). q_p (or Fq'/Fv') is the photochemical quenching coefficient under actinic
 227 light estimated as the difference of apparent PSII photochemical efficiencies measured
 228 in 'light' and 'dark' chambers of the FRRf (see Table 2), following the procedure of
 229 Suggett et al. (2006a,b, 2011). The factor 0.0243 converts seconds to hours, μmol
 230 quanta to mol quanta, \AA^2 to m^2 and mol Chl-a to mg Chl-a. Following Kolber and
 231 Falkowski (1993), n_{PSII} was assumed to be a constant i.e. 0.002 mol RCII mol
 232 Chl-a⁻¹ since the phytoplankton populations we examined were mostly dominated by
 233 eukaryotes, mainly diatoms, *Skeletonema spp.*, and dinoflagellates, *Ceratium furca*
 234 (Shibata et al., 2010). It may be noted that assigning a constant value to n_{PSII} has
 235 been recognised as a potential source of error to ETR determinations (e.g. Suggett et
 236

237 al. 2004, 2009b, 2011), which we discuss later.

238 Previous work has shown that differences between ^{13}C -based CO_2 uptake rates
239 and FRRf-based *in situ* ETRs appear to predominantly reflect differences in light
240 fields in the two measurements (Tripathy et al., 2010). Thus, in order to account for
241 high frequency fluctuations in the incident irradiance field between the FRRf casts
242 (every 2 hours), a simple modelling approach was utilized to estimate daily-integrated
243 ETR, which was subsequently used for comparison against daily-integrated ^{13}C
244 uptake rates. Our approach is analogous to those previously used for determining
245 daily ETR (Suggett et al. 2001, Smyth et al. 2004, Mino et al., 2014).

246 We determined ETR versus PAR relationships for each light depth corresponding
247 to the 6 sample depths of the ^{13}C uptake experiments (100, 50, 25, 10, 5 and 1% of the
248 $\text{PAR}(0^-)$. Incident PAR logged continuously at the surface was multiplied by the factor
249 0.9 to account for physical properties at the air-sea interface which reduce PAR above
250 the water surface, $\text{PAR}(0^+)$, relative to that just beneath the surface, and $\text{PAR}(0^-)$
251 (Marra, 2015). Underwater PAR at the % light depth of interest (x%) was then
252 determined as $\text{PAR}(0^-)(t) \times (x/100)$.

253 ETR versus instantaneous PAR relationships were constructed for each cruise
254 using data from the FRRf time series casts, and the model of Jassby and Platt (1976),
255 to determine α and ETR_{max} which are two fitted parameters describing the initial
256 slope of the ETR vs PAR curve and the maximum electron transfer rate,

257

$$258 \quad \text{ETR}(z, t) = \text{ETR}_{max} \times \tanh\left(\frac{\alpha \text{PAR}(z, t)}{\text{ETR}_{max}}\right), \quad (4)$$

259 .

260 With knowledge of α and ETR_{max} , we were then able to retrieve the ETR for
261 any given value of PAR from the continuous underwater light field and thus determine

262 the daily integrated ETR at specific depth, according to following equation:

263

$$264 \quad ETR_daily(z) = \int_{t_1}^{t_2} ETR(z, t) dt, \quad (5)$$

265

266 where the period between t_1 and t_2 represent the time in hours from dawn to dusk.

267 Finally, the electron requirement of carbon uptake K_C ($\text{mol e}^- (\text{mol C})^{-1}$) was

268 calculated as,

$$269 \quad K_C = ETR_daily / P^B \times 12, \quad (6)$$

270

271 where ETR_daily ($\text{mol e}^- \text{mg Chl-a}^{-1} \text{d}^{-1}$) and P^B ($\text{mgC mg Chl-a}^{-1} \text{d}^{-1}$) were the

272 daily-integrated ETR and daily-integrated carbon assimilation per unit Chl-a, and the

273 factor 12 converts g C to mol C.

274 In addition to determining ETR, FRRf casts were also used to determine a proxy

275 for non-photochemical quenching (NPQ), so as to verify recent observations as to

276 whether K_C (under conditions where n_{PSII} is unknown and hence an assumed

277 constant) was closely related to the extent of NPQ (Schuback et al., 2015).

278 Specifically, the photochemical efficiency measured in the dark chamber for any

279 given depth, $F_v/F_m(z)$, was normalised to the maximum photochemical efficiency

280 measured throughout the water column (Suggett et al., 2006b; 2011); this maximum

281 typically occurs at low light depths where the influence of surface-driven NPQ is

282 eliminated and hence corresponds to the dark-acclimated maximum efficiency

283 $(F_v/F_m)_M$ (Smyth et al., 2004; Suggett et al. 2011). Thus our NPQ Proxy (NPQ*) (z)

284 = $1 - [F_v/F_m(z) / (F_v/F_m)_M]$. In order to compare NPQ* with the daily K_C for each

285 light depth, we determined the daily mean NPQ* (z) from across all casts throughout

286 the day.

287 **2.5 Statistical analyses**

288 Spearman Rank Order Correlation analysis was used to evaluate the key
289 environmental variables that may be related with K_C . Correlations were considered
290 significant when p was < 0.05 . Stepwise Multiple Regression (SMR) was then utilized
291 to generate the model for estimating the dependent variable (i.e. K_C) from the various
292 independent variables (i.e. environmental factors). p -values (< 0.05) from the SMR
293 were used as the criterion for choosing environmental variables entered into the K_C
294 prediction model. The Spearman Rank Order Correlation Analyses and SMR were
295 performed using open source statistical software R version 3.1.0 (R Core Team 2014).
296

297 **3. Results**

298 **3.1 Physicochemical characteristics and phytoplankton biomass**

299 Cruises were conducted in fall November 2008 and 2009 and in spring May 2008
300 and 2010. Water column properties during the two seasons when these cruises were
301 undertaken were quite different (Table 1, Fig. 2). Vertical profiles of PAR, temperature,
302 salinity, and density plotted for each cruise (Fig. 2), revealed higher daily PAR, lower
303 temperature and lower surface salinity during May as compared to Nov (Fig. 2a-c).
304 Mixed layer depth, defined as depth at which the density change from the surface
305 density was 0.03 kg m^{-3} (Shibata et al., 2010), generally showed deeper mixing in
306 Nov than May (Table 1, Fig. 2d). Surface PAR values were higher in May; however,
307 the average euphotic depth, Z_{eu} , and average diffuse attenuation coefficient, K_d ,
308 generally similar for the two seasons (May, $13.5 \pm 1.9 \text{ m}$, $0.33 \pm 0.04 \text{ m}^{-1}$; November,
309 $16.3 \pm 3.2 \text{ m}$, $0.29 \pm 0.07 \text{ m}^{-1}$). Thus similar vertical attenuation of light in the water
310 column was evident for both seasons.

311 Average inorganic nutrient concentrations, $\text{NO}_3^- + \text{NO}_2^-$, PO_4^{3-} and SiO_2 were
312 13.8 ± 0.6 , 1.34 ± 0.04 and $57 \pm 3 \text{ } \mu\text{M}$, respectively in November, and generally
313 constant throughout the mixed water column (Fig. 3a,b,c) as expected given the
314 uniform profiles of density with depth (Fig. 2d). $\text{NO}_3^- + \text{NO}_2^-$ and PO_4^{3-} concentrations
315 were lower in May than for November (0.4 ± 0.1 and $0.16 \pm 0.06 \text{ } \mu\text{M}$). The ratio of
316 $\text{NO}_3^- + \text{NO}_2^-$ and PO_4^{3-} concentrations (N/P ratio) was < 5 across all depths in May but
317 increased significantly in November, and with mean value of 10.2 (Fig. 3d). Chl-a
318 concentrations were generally similar across all depths and consistently lower in May
319 ($1.18\text{-}3.87 \text{ mg m}^{-3}$) than in November ($1.92\text{-}6.61 \text{ mg m}^{-3}$) except in 15 Nov 2009 (Fig.
320 3e).

321 3.2 Variability of FRRf parameters and ETR

322 Time series profiles of FRRf derived photo-physiological parameters obtained
323 throughout the day on 21 May 2010 (hereafter referred to as May) and 8 November
324 2008 (hereafter November) are presented in Figs. 4a-d and Figs. 4e-h respectively.
325 The FRRf data from these two cruises are taken as representative profiles associated
326 with the differences in hydrographical and climatic conditions during the two seasons
327 (shallow vs. deep mixed layer; high vs. low daily PAR).

328 In May, the ratio for σ_{PSII} spectral weighting, $\bar{a}^{chl}(in\ situ) : \bar{a}^{chl}(FRR)$
329 varied from 0.60 at surface to 0.49 at 1% PAR depth, with mean 0.54 ± 0.04 (Fig. 5).
330 Considering this small vertical variation, we used the mean value for correcting σ_{PSII}
331 across all depths. σ_{PSII}^{abs} values plotted after this correction varied over depth and time,
332 from 248 to 386 $\text{\AA}^2 \text{ quanta}^{-1}$, and a mean value of $334 \pm 30 \text{\AA}^2 \text{ quanta}^{-1}$ (Fig. 4a).
333 σ_{PSII}^{abs} values increased with depth, from the surface to ca. 8 m depth (10% surface
334 PAR depth), at which point they appeared to plateau and attain relatively constant
335 values in the deeper layer. Over the course of the day, maximum and minimum values
336 of near surface σ_{PSII}^{abs} were observed at 6:00 and 10:00 cast, respectively. The mean
337 σ_{PSII}^{abs} of the upper layer (from the surface to 10% PAR depth) was calculated to be
338 $303 \pm 35 \text{\AA}^2 \text{ quanta}^{-1}$. In November, $\bar{a}^{chl}(in\ situ) : \bar{a}^{chl}(FRR)$ was slightly larger
339 than that observed in May, with a mean value 0.64 ± 0.03 (0.69 to 0.58; Fig. 5). After
340 applying this spectral correction, σ_{PSII}^{abs} varied from only 296 to 403 $\text{\AA}^2 \text{ quanta}^{-1}$, and
341 with a higher mean value of $360 \pm 23 \text{\AA}^2 \text{ quanta}^{-1}$ indicative of low light acclimation
342 (Moore et al., 2006). As observed in May, σ_{PSII}^{abs} values increased with depth, but the
343 difference in σ_{PSII}^{abs} values throughout the water column was smaller in November
344 than May (Fig. 4a, e).

345 Values of the maximum photochemical efficiency, F_v/F_m (dimensionless) were
346 relatively similar for both cruises, ranging 0.31-0.52 in May and 0.36-0.54 in
347 November, respectively. Below 10% PAR depth (8m) variation of F_v/F_m over the
348 course of the day was generally low compared to that at the surface. The reduced
349 values of F_v/F_m in the upper layer during the day can be attributed to the NPQ
350 (Suggett et al., 2011) (Fig. 4b, f). Values for q_p during both cruises were close to the
351 maximum of 1 at depths > 10 m and lowest at the surface (Fig. 4c, g). Surface values
352 of q_p were lower (0.5) in May as compared to 0.64 in November. These differences
353 in photochemical quenching reflect differences in surface irradiances in May and in
354 November.

355 Changes in ETR over the course of the day reflected the variations in ambient
356 PAR, σ_{PSII}^{abs} and q_p (Table 3). The large changes in ETR values at depths < 10 m
357 during the May cruise are from the larger differences in σ_{PSII}^{abs} and q_p during May.
358 Vertical profiles of ETR values maintained a consistent trend with depth (and at
359 depths shallower than 10 m differed appreciably over the course of the day)
360 throughout, but maximum values in May were almost twice as that in November (Fig.
361 4d, h).

362 Light response models applied to ETR versus PAR scatter plots for all seven
363 cruises are shown in Figs. 6a-g, and the fitted parameters presented in Table 4; these
364 parameters were then utilized to derive the daily ETR based on Eq. 5 (Fig. 7a). Daily
365 ETR calculated from all pooled data and plotted versus optical depths varied
366 considerably, from 0.5 to 115.7 mmol e⁻ (mg Chl-a)⁻¹ d⁻¹, resulting in an average (±
367 standard deviation) of 24.8 ± 27.5 mmol e⁻ (mg Chl-a)⁻¹ d⁻¹ (Fig. 7a). The daily ETR
368 of upper lit layer (> 10% PAR depth) averaged 60.4 ± 29.9 mmol e⁻ (mg Chl-a)⁻¹ d⁻¹,
369 n=18) in May and was almost twice of the ETR observed in November, 33.6 ± 17.3

370 $\text{mmol e}^- (\text{mg Chl-a})^{-1} \text{d}^{-1}$, $n=24$). For all cruises, ETR increased progressively from
371 deeper to shallower optical depths. The relationship between daily ETR and PAR
372 plotted in Fig. 7b shows a significant correlation ($r^2=0.98$, $n=42$, $p < 0.001$) between
373 these two parameters, with fitted maximum ETR of ca. $120 \text{ mmol e}^- (\text{mg Chl-a})^{-1} \text{d}^{-1}$.

374 **3.3 Carbon uptake rates**

375 Daily Chl-a normalised carbon uptake rates (P^B) varied from 2.95 to 69.9 mgC mg
376 $\text{Chl-a}^{-1} \text{d}^{-1}$, with an average value of $31.65 \pm 17.09 \text{ mgC mg Chl-a}^{-1} \text{d}^{-1}$. Profiles of P^B
377 plotted in Fig. 8a showed signs of photoinhibition during several cruises, which was
378 not consistent with the profiles of daily ETRs in which no subsurface maximum was
379 observed. Differences in the highest and lowest P^B for the two seasons were also small
380 (69.9 and $2.95 \text{ mgC mg Chl-a}^{-1} \text{d}^{-1}$ in May; 58.47 and $3.36 \text{ mgC mg Chl-a}^{-1} \text{d}^{-1}$ in
381 Nov.). Amongst all cruises, the mean value of P^B_{opt} (defined as the maximum P^B
382 within the water column) in May, $62.7 \text{ mgC mg Chl-a}^{-1} \text{d}^{-1}$, was larger than that for
383 November, $47.0 \text{ mgC mg Chl-a}^{-1} \text{d}^{-1}$.

384 P^B values were plotted against corresponding ETR values (Fig. 8b) to examine the
385 extent with which water column primary productivity values can be explained by ETR.
386 Overall, values of P^B and ETR generally co-varied well for lower ETR values.
387 However, when daily ETR values exceeded ca. $20 \text{ mmol e}^- [\text{mgChl-a}]^{-1} \text{d}^{-1}$, relatively
388 large variations between P^B and ETR were observed suggesting that beyond a certain
389 threshold, measurements of daily ETR cannot adequately reflect carbon fixation rates.

390 **3.4 Variability of the quantum requirement (K_C)**

391 One of our goals was to examine whether the breakdown of the relationship
392 between ETR and P^B could be explained by changes in values of K_C over the course
393 of the day or with depth. Values of K_C determined from corresponding values of daily

394 ETR and C-uptake (Eq. 6) exhibited a range of 2.3-26.6 mol e⁻ (mol C)⁻¹ in May
395 (mean ± standard deviation: 13.2 ± 17.9 mol e⁻ (mol C)⁻¹) and 1.2-19.3 mol e⁻ (mol
396 C)⁻¹ in November (mean: 5.7 ± 4.7 mol e⁻ (mol C)⁻¹), and were always higher for
397 surface than deep waters (Fig. 9a).

398 K_C values at higher optical depths (i.e. 1% and 5% of PAR at the surface, roughly
399 < 3 mol quanta m⁻² d⁻¹) recorded during our study were lower than the theoretical
400 minimum value of 4 mol e⁻ (mol C)⁻¹. Since the shallower optical depths (shallower
401 than the 10% PAR depth) contribute to ~80% of water column integrated NPP (IPP) in
402 our study area, we therefore excluded K_C values where daily PAR < 3 mol quanta m⁻²
403 d⁻¹ from further analysis. With the exclusion of this data, K_C varied from 4.1 to 26.6
404 mol e⁻ (mol C)⁻¹ (Fig. 9b).

405 Variability of K_C was finally examined in the context of the prevailing
406 environmental variables using Spearman Rank Order Correlation analysis. SMR
407 analysis was further used to establish the relationship between the key environmental
408 regulators of K_C (*p* < 0.05). These analyses revealed that K_C was highly related to
409 incident irradiance (*r*² = 0.941, *p* < 0.001; Table 5). SMR provided the best model fit
410 for predicting K_C on the basis of PAR (Fig. 9c). Using the same analytical steps we
411 were further observed a significant correlation between K_C and non-photochemical
412 quenching (NPQ*) (Fig. 9d).

413 **4. Discussion**

414 Previous FRRf-based observations in Araiike Bay study (Tripathy et al. 2010)
415 relied on a constant value (= 4) for K_C for estimating daily photosynthetic rates and
416 subsequent comparison with carbon uptake measurements from ¹³C incubation
417 experiments; this work demonstrated reasonably good agreement between FRRf and

418 C uptake based primary productivity when ambient light intensities were low.
419 However, Tripathy et al. (2010) considered that the FRRf method overestimated daily
420 net primary production two to three fold in particular at the surface (i.e. under high
421 light conditions) and hypothesised that the assumed constant K_C likely led to
422 significant errors in FRRf-based primary productivity estimates, especially when light
423 intensities are high. Our current study has directly addressed and confirmed this
424 hypothesis. Specifically, we show for the first time that the electron requirement for
425 net carbon fixation (K_C) varies in a predictable manner with the daily irradiance. These
426 observations are consistent with recent reports suggesting that K_C is highly variable
427 but can potentially be predicted from knowledge of key environmental variables
428 (Lawrenz et al., 2013). In the following sections, we consider our results that net
429 primary production can be estimated by FRRf directly when light-dependent
430 variability in K_C is accounted for in the eutrophic enclosed bay, Ariake Bay.

431 **4.1 Reconciliation of ETRs and ^{13}C uptake rates**

432 Values of K_C from the upper layer in our study varied from 4.1 to 26.6 mol e^-
433 (mol C) $^{-1}$, with a mean and standard deviation of 9.6 ± 5.4 mol e^- (mol C) $^{-1}$. These
434 values are within the range of 1.15-54.2 (mean: 10.9 ± 6.91) mol e^- (mol C) $^{-1}$ global
435 observations of K_C (Lawrenz et al. 2013), which were derived in large part from
436 short-term incubations, typically 1-4 h. In principle K_C should increase as C-uptake
437 increasingly transitions from gross to net (e.g. Halsey et al. 2013); thus, the general
438 agreement of our mean K_C values to those previously measured elsewhere (Lawrenz
439 et al. 2013) would suggest net carbon uptake rates may already be attained after
440 relatively short incubations. However, our higher values are generally more consistent
441 with those measured under conditions of nutrient stress and/or limitation, which

442 elevates K_C (Lawrenz et al. 2013). Since our study did not appear to be under the
443 influence of nutrient stress the higher values we observed could indeed reflect
444 ‘inflation’ from a prolonged incubation period of net carbon uptake. Unfortunately it
445 is presently impossible to reconcile the confounding influence of incubation length
446 and the broad variation that appears evident for K_C across studies/regions, and thus
447 warrants further attention.

448 In previous FRRf-based studies (Corno et al. 2006; Melrose et al. 2006; Suggett
449 et al. 2006a; 2009a), values for K_C lower than the theoretical minimum (i.e. 4 mol e^-
450 $(\text{mol C})^{-1}$) have often been measured but still hard to resolve. In fact the long
451 incubation times during SIS experiments generally yield CO_2 uptake closer to net
452 primary production (Marra, 2009), and thus increases the expected minimum value for
453 K_C . Specifically, net CO_2 uptake rates are typically lower than gross uptake rates by a
454 factor of ca. 6-55%, depending on respiration loss of different phytoplankton species
455 and their growth phase (López-Sandoval et al., 2014). This suggests the minimum
456 value for K_C under net carbon uptake conditions should ultimately be considered
457 higher than $4 \text{ mol e}^- [\text{mol C}]^{-1}$. When all sample depths were considered in our study,
458 values of $K_C < 4 \text{ mol e}^- (\text{mol C})^{-1}$ contributed ca. 40% of the total number of samples.

459 K_C values below the ‘theoretical minimum’ are generally considered an
460 overestimation of carbon uptake and/or underestimation of ETR (Suggett et al., 2009a;
461 Lawrenz et al. 2013). Therefore we further analysed our data set excluding values of
462 K_C from deeper waters (where values $< 4 \text{ mol e}^- (\text{mol C})^{-1}$, were observed) with the
463 justification that net primary productivity at these depths generally contribute a very
464 small fraction of integrated euphotic water column net primary production. In this
465 way we were able to generate a clear positive correlation between upper layer K_C and
466 daily irradiance (and with the intercept of the regression close to the theoretical

467 minimum for K_C . Even so, greater understanding of why values for K_C often drift < a
468 value of 4 requires more focused study in future. Several potential variables can
469 contribute to potentially erroneous ETRs (see Suggett et al. 2009a). Our ETR
470 calculation (Eq. 3) assumed a ‘standard’ n_{PSII} value of $0.002 \text{ mol RCII [mol chl]}^{-1}$
471 for eukaryotes (Kolber & Falkowski 1993) since phytoplankton in this region were
472 mainly dominated by diatoms and dinoflagellates during the two study seasons
473 (Shibata et al., 2010; Tripathy et al., 2010). Suggett et al. (2011) reported a mean
474 n_{PSII} of $0.00188 \pm 0.00002 \text{ mol RCII [mol Chl-a]}^{-1}$ for 2 and 7 species of
475 dinoflagellates and diatoms under various growth conditions. Thus, the observations
476 of Suggett et al. (2011) suggest that our use of constant n_{PSII} would potentially lead
477 to very small errors in the estimation of K_C . Our values for σ_{PSII} was spectrally
478 corrected, and ‘sample blanks’ not required based on our approach to estimate qP
479 (Suggett et al. 2006a, b) and therefore also unlikely contributed a major course of
480 error in our ETRs.

481 **4.2 Decoupling of ETR from carbon uptake**

482 Our study shows that the decoupling of ETR from 24-h net C-uptake was most
483 significant at higher irradiances, a finding that is consistent with the observations of
484 Schuback et al. (2015) across a broad range of biogeographic areas, with shorter
485 incubation time (3-4 hours). Light-dependency of K_C has been reported for
486 microalgae, and generally higher values of K_C have been recorded under
487 light-saturated conditions (Suggett et al., 2008; Brading et al., 2013; Hancke et al.,
488 2015). High ETRs can be sustained through up-regulation of alternative electron flow
489 pathways (Prasil et al., 1996; McDonald et al., 2011) and/or light-dependent O_2
490 consuming processes (Suggett et al. 2008; Mackey et al. 2008). Such processes, which
491 include Mehler Ascorbate Peroxidase and plastoquinone terminal oxidase (PTOX)
492 activity, act to balance the availability of light energy with the amount of reductant in
493 cells when rates of linear electron transport exceed the capacity for CO_2 assimilation
494 (e.g. Cardol et al., 2011).

495 The relationship between K_C and light availability may further explain why we
496 also observed co-variability of K_C with the inherent NPQ capacity (Fig. 8d),
497 observations that are again consistent with those of Schuback et al (2015) but from
498 iron-limited waters. Sustained electron transport activity under high light and hence
499 ΔpH -triggering NPQ has been reported previously for both diatoms and
500 dinoflagellates (Ott et al., 1999; Lavaud et al., 2004), i.e. groups that generally
501 dominate Ariake Bay (see above). It is known that NPQ shows predictable
502 light-dependency dynamics as a result of energy-dependent quenching in the light
503 harvesting antennae (e.g. Serodio & Lavaud 2011). So clearly excess light would also
504 increase the NPQ expression and explain the linear relation between K_C and NPQ*
505 observed here.

506 Despite the observed light regulation of K_C in our study, we cannot fully exclude
507 the potential effects of other factors that could co-vary with light availability (Table 5).
508 Specifically, the highest values of K_C corresponded to the lowest $\text{NO}_3^- + \text{NO}_2^-$ and
509 PO_4^{3-} concentrations found in the upper stratified waters in May, i.e. when light
510 intensity was also high. Previous comparison studies between ETR and C-uptake
511 clearly show variability of K_C was correlated to that of temperature and/or inorganic
512 nutrients rather than light availability in some biogeographic regions (see Lawrenz et
513 al, 2013). Such environmental factors also drive large changes in both phytoplankton
514 community structure and physiological status, both of which can drive variance in K_C
515 (Debes et al., 2008; Kromkamp et al., 2008; Suggett et al., 2006b, 2009a; Robinson et
516 al., 2014).

517 To evaluate for potential effects of nutrient concentrations on K_C , we compared
518 values for K_C from two cruises (28 May, 2008 vs. 08 Nov., 2008) from days with
519 similar daily light intensity (ca. 19 vs. 15 mol quanta $\text{m}^{-2} \text{d}^{-1}$) but with large
520 differences in nutrient concentrations (average $\text{NO}_3^- + \text{NO}_2^-$: 0.4 vs. 14.5 μM ; PO_4^{3-} :
521 0.22 vs. 1.18 μM). Despite differences in nutrient conditions for these two cruises, the
522 K_C values were similar i.e. $7.9 \pm 2.4 \text{ mol e}^- (\text{mol C})^{-1}$ for May 28, 2008 as compared
523 to $8.5 \pm 1.8 \text{ mol e}^- (\text{mol C})^{-1}$ for November 06 2008. Thus, any influence upon K_C
524 from nutrients is likely secondary to that from light in our study area. Nitrogen is
525 considered to be limited in May (N/P: ca. 2.6), and appeared relatively low
526 concentration in water column (0.16 – 0.84 μM); however it should be noted that our
527 nitrogen data does not include ammonium-nitrogen (NH_4^+ -N), which may account for
528 ca. 30% of the total DIN amount (Tabata et al., 2015). Thus it is unlikely nutrient
529 stress is significant even in May in Araiike Bay. Comparable measurements of F_v/F_m
530 for May and November further suggest a lack of major nutrient stress (but see Suggett

531 et al. 2009b). This absence of nutrient stress probably explains why we were not able
532 to observe a robust relationship between nutrients and K_C .

533 Our results suggest that in the Ariake Bay, where land-derived nutrient inputs are
534 large, phytoplankton rarely experience nutrient starvation. Under these conditions,
535 incident PAR appears the main factor responsible for the variability in K_C . This result
536 contrasts with that of Lawrenz et al. (2013), where variance of K_C appeared more
537 commonly governed by that of temperature, nutrients and/or light attenuation.
538 However, this may reflect a lower range of temperature and nutrients (and Chl-a)
539 observed across our dataset relative to the variance in K_C measured. Furthermore,
540 Lawrenz et al. (2013) do not consider absolute PAR as a variable only K_d and optical
541 depth within their meta-analysis, thus it is impossible to verify how local light
542 conditions at the time of sampling may potentially have further contributed to the
543 variability they observed. We have shown that K_C can be estimated by a simple light
544 dependence from easily measurable irradiance in this region. Whether the resulting
545 regression may perhaps be applicable to the other locations with similar
546 environmental characteristics (i.e. absence of nutrient limitations and/or homogenous
547 distribution of phytoplankton groups) remains to be seen.

548 **4.3 Recommendations for future *in situ* application**

549 The final objective of this study was to estimate NPP from FRRf directly. We
550 have demonstrated that the key factor (K_C) for converting FRRf based ETR values to
551 NPP can be derived from light intensity. The ETR vs. PAR relationship from FRRf
552 deployments around noon when instantaneous PAR was highest, combined with
553 continual surface PAR records, could be applied for regional estimates of NPP
554 estimation. The advantage here of active fluorometry is that a large amount of *in situ*

555 data covering broad spatial (and temporal) scales can be acquired, which in turn can
556 further be utilized for validation of NPP derived from satellite data (as in Behrenfeld
557 et al., 1997; Kameda and Ishizaka, 2005; Hirawake et al., 2012). Before this becomes
558 a real possibility, several issues must first be addressed:

559 Firstly, the need for additional spectrally-resolved absorption and in-water
560 irradiance measurements to spectrally correct σ_{PSII} . Suggett et al. (2006b) previously
561 overcame this problem for FRRf data collected across many water types using an
562 algorithm between the spectral correction factor and optical depth. For our study, we
563 observed reactively little variability of the correcting factor for σ_{PSII} , and therefore
564 for practical reasons used a constant value for each season (i.e. 0.54 in spring/summer
565 and 0.64 in fall/winter; and thus an overall average value (i.e. 0.6) for σ_{PSII} . This
566 approach clearly simplifies broad-scale deployment of the FRRf but requires further
567 verification.

568 Secondly, it is not possible to measure n_{PSII} in most natural samples but some
569 knowledge of the RCII concentration can be obtained indirectly (Suggett et al., 2011)
570 or through new algorithms for calibrating [RCII] from the FRRf parameters
571 themselves (see Oxborough et al. 2012). In the semi-enclosed water region as Ariake
572 Bay or estuary, where nutrient concentrations remain relatively high and the
573 phytoplankton community dominated by micro-phytoplankton, n_{PSII} was considered
574 to remain at a constant value of $0.002 \text{ mol RCII (mol Chl-}a\text{)}^{-1}$ but again ultimately
575 requires further verification. Furthermore, using algorithms for K_C not dependant on
576 knowledge of n_{PSII} (Schuback et al. 2015) provide a means to overcome this
577 limitation.

578 As a further test of the validity of our FRRf-based approach for estimating carbon
579 uptake, we applied the procedures described above ($\text{ETR}_{\text{daily}} \times K_C$) to estimate NPP

580 from FRRf casts deployed in 2007 winter, and subsequently compared these FRRf
581 derived NPP values with those obtained from daily ^{13}C -incubation experiments. All
582 methods were the same as for 2008-2010 cruises but with no underwater spectral
583 irradiance and/or phytoplankton absorption data was not available (so constant
584 spectral correction factor of 0.64 was applied). Even so, estimation of P^B using the
585 FRRf data alone gave good agreement with the measured ^{13}C uptake, albeit with a
586 slight underestimation under conditions of high carbon uptake (Fig. 10a). We also
587 tested this FRRf-based P^B approach for estimating daily- and depth-integrated NPP
588 for the entire dataset (2007 – 2010) (Fig. 10b). Here a strong correlation was observed
589 between estimated and measured P^B and IPP both were found, and RMSE for the P^B
590 and IPP were $12.6 \text{ mgC (mgChl-a)}^{-1} \text{ d}^{-1}$ and $271.8 \text{ mgC m}^{-2} \text{ d}^{-1}$, respectively.

591 Current satellite NPP models are still considered to perform poorly in Case-2
592 waters (Saba et al., 2011) and need to be modified or localized for specific regions,
593 which in turn is dependent upon how much data is locally available (Tripathy et al.,
594 2012); thus, our FRRf-based NPP approach may provide a major step towards this
595 currently limitation. For example, popular models such as the Vertically Generalized
596 Production Model (VGPM) locally require algorithms that can improve upon the
597 generally applied optimal rate of productivity (i.e. P_{opt}^B) (Kameda and Ishizaka, 2005)
598 and thus definitely benefit from the high data volume afforded through FRRf. Thus,
599 whilst our present study has provided a first look at an improved FRRf method for
600 estimating NPP in a eutrophic embayment, we suggest that it lays the foundation for a
601 broader scale complementary approach to yield carbon-based NPP measurements for
602 further satellite model validation and/or improvement.

603

604

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615 **Reference**

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778 of the Changjiang River on the light absorption properties of phytoplankton from
779 the East China Sea. *Biogeosciences* 11: 1759-1773.
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782 **Lists of tables**

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785 **Table 1** Details of the sampling and environmental characteristics during the
786 sampling.

No.	Sampling date	Sampling periods	Daily PAR(0 ⁺) (mol quanta m ⁻² d ⁻¹)	Z _{eu} (1% surface PAR, m)	Mixed layer depth (m)
1	28 May, 2008	6:00 – 18:00	19.07	15.0	5
2	08 Nov., 2008	8:00 – 16:00	14.99	20.0	19
3	15 Nov., 2008	8:00 – 16:00	14.58	11.0	17
4	08 Nov., 2009	6:00 – 16:00	33.68	16.8	6
5	15 Nov., 2009	6:00 – 16:00	24.19	17.0	12
6	14 May, 2010	6:00 – 18:00	68.18	11.0	7
7	21 May, 2010	6:00 – 18:00	53.36	15.2	8

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Table 2 Definitions of photosynthetic parameters. (^a Referred to Suggett et al., 2006a,b)

Parameter	Definition
E	Instantaneous irradiance ($\mu\text{mol quanta m}^{-2} \text{ s}^{-1}$)
F_o	Minimum fluorescence yield in dark chamber (arbitrary units: a.u.)
F_m	Maximum fluorescence yield in dark chamber (a.u.)
F_v/F_m	Potential photochemical efficiency of open reaction centers [$=(F_m - F_o)/F_m$] (dimensionless)
F'	Steady-state fluorescence yields in light chamber (a.u.)
F_m'	Maximum fluorescence yield in light chamber (a.u.)
F_q'/F_m'	Photochemical efficiency of PSII under actinic light
$q_p (F_q'/F_v')$	Photochemical quenching coefficient, as the difference in the apparent PSII photochemical efficiency between FRRf light and dark chamber quasi-simultaneously, = $\left[\frac{(F_m' - F')/F_m' \text{ light chamber}}{(F_m - F_o)/F_m \text{ dark chamber}} \right]^{\text{[a]}}$, (dimensionless)
n_{PSII}	Photosynthetic unit size of PSII (=0.002) ($\text{mol RCII} (\text{mol Chla})^{-1}$)
σ_{PSII}	Effective absorption cross section of PSII in dark chamber ($\text{\AA}^2 \text{ quanta}^{-1}$)
σ_{PSII}'	Effective absorption cross section of PSII in light chamber ($\text{\AA}^2 \text{ quanta}^{-1}$)
σ_{PSII}^{abs}	Spectral corrected effective absorption cross section of PSII ($\text{\AA}^2 \text{ quanta}^{-1}$)
ETR	Rate of electron transport through PSII ($\text{mmol e}^- (\text{mg Chl-a})^{-1} \text{ h}^{-1}$)
K_C	Electron requirement for carbon fixation ($\text{mol e}^- (\text{mol C})^{-1}$)

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798 **Table 3** Summary of PAR, σ_{PSII}^{abs} and q_p ranges in two example cruises

Cruise time	instantaneous PAR (surface) ($\mu\text{mol quanta m}^{-2} \text{ s}^{-1}$)	σ_{PSII}^{abs} ($\text{\AA}^2 \text{ quanta}^{-1}$)	q_p (surface) (dimensionless)
08 Nov., 2008 (Nov.)	45.9 - 665.1	296 - 403	0.64 - 0.96
21 May, 2010 (May)	56.7 - 1321.8	248 - 386	0.50 - 0.94

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Table 4 The estimated ETR-irradiance parameters for each cruise.

	May 28, 2008	Nov. 08 2008	Nov. 15 2008	Nov. 08 2009	Nov. 15 2009	May 14 2010	May 21 2010
α	0.016 (0.001)	0.015 (0.001)	0.014 (0.001)	0.012 (0.001)	0.011 (0.001)	0.011 (0.001)	0.010 (0.001)
ETR_{max}	8.42 (0.19)	7.66 (0.12)	12.11 (0.58)	11.43 (0.33)	15.64 (0.24)	10.64 (0.31)	11.06 (0.21)

804 Initial slope of the ETR-PAR curve (α) and the estimated maximum ETR value
805 (ETR_{max}) with fitting standard error

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810 **Table 5** Spearman correlation coefficients for correlations between daily K_C and
811 environmental variables

	PAR	Temp.	Salinity	Chl-a	$NO_3^-+NO_2^-$	PO_4^{3-}	SiO_2
K_C	0.941** n=20	-0.185 n=20	-0.200 n=20	0.065 n=20	-0.211 n=20	-0.298 n=20	-0.0484 n=20

812 ** indicates significance of the correlation at the 0.01 significant level.

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814 **Lists of figures**

815

816 **Fig. 1** Location of Ariake Bay and sampling station.

817 **Fig. 2** Representative vertical profiles **(a)** daily PAR ($\text{mol quanta m}^{-2} \text{d}^{-1}$); **(b)**
818 temperature ($^{\circ}\text{C}$). Arrows indicate Z_{eu} depth of each cruise; **(c)** salinity and **(d)**
819 potential density from pre-dawn CTD cast of each cruise. The two profiles with solid
820 lines were typical water characteristics in May and November; specifically cruises
821 from 21 May 2010 and 8 November 2008.

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823 **Fig. 3** Representative vertical profiles **(a)** $\text{NO}_3^- + \text{NO}_2^-$ (μM); **(b)** PO_4^{3-} (μM); **(c)** SiO_2
824 (μM), **(d)** N/P ratio (dimensionless) and **(e)** Chl-a (mg m^{-3}).

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826 **Fig. 4** Time series profiles of FRRf-based data $\sigma_{\text{PSII}}^{\text{abs}}$ **(a, e)**, F_v/F_m (dimensionless; **b,**
827 **f**), q_p (dimensionless; **c, g**) and ETR ($\text{mmol e}^- [\text{mgChl-a}]^{-1} \text{h}^{-1}$; **d, h**). Upper and
828 lower panels are for cruise data of 21 May 2010 (a-d) and 8 November 2008 (e-h)

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830 **Fig.5** Vertical profiles of the σ_{PSII} correction factor (dimensionless) in 21 May 2010
831 and 8 November 2008 cruise. Absorption data was collected from the same cast as
832 for the water samples for ^{13}C uptake experimentation; underwater spectral irradiance
833 was measured at local noon

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835 **Fig. 6** ETR($\text{mmol e}^- [\text{mgChl-a}]^{-1} \text{h}^{-1}$) – PAR($\mu\text{mol quanta m}^{-2} \text{s}^{-1}$) fitting results for
836 cruises on **a)** 28 May, 2008; **b)** 08 Nov., 2008; **c)** 15 Nov., 2008; **d)** 08 Nov., 2009; **e)**
837 15 Nov., 2009; **f)** 14 May, 2010; **g)** 21 May, 2010. Scatter data is the measured FRRf
838 based ETRs from all time series casts and the dotted line is the model fit. Fitting
839 statistics for each curve are summarized in Table 4.

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841 **Fig. 7 (a)** Daily-integrated ETR ($\text{mmol e}^- [\text{mgChl-a}]^{-1} \text{d}^{-1}$) profiles plotted against
842 optical depth for each cruise; and **(b)** plots of daily-integrated ETR (mmol e^-
843 $[\text{mgChl-a}]^{-1} \text{d}^{-1}$) against daily-integrated PAR ($\text{mol quanta m}^{-2} \text{d}^{-1}$). The dotted lines
844 are the fitted tanh function.

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847 **Fig. 8 (a)** ^{13}C -uptake based determinations of Chla-specific carbon uptake (P^B , mgC
848 $\text{mg Chl-a}^{-1} \text{d}^{-1}$) plotted against optical depth for each cruise; and **(b)** Scatter plots of
849 P^B ($\text{mmol C} [\text{mgChl-a}]^{-1} \text{d}^{-1}$) against corresponding measurements of daily-integrated
850 ETR ($\text{mmol e}^- [\text{mgChl-a}]^{-1} \text{d}^{-1}$). Dotted line represents where ETR is 20 mmol e^-
851 $[\text{mgChl-a}]^{-1} \text{d}^{-1}$, i.e the point at which the linear correlation between ETR and P^B
852 appears to break down.

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855 **Fig. 9** Profiles of **(a)** all K_C values ($\text{mol e}^- (\text{mol C})^{-1}$) and **(b)** excluding data where K_C
856 $< 4 \text{ mol e}^- (\text{mol C})^{-1}$; **(c,d)** Scatter plot of the relationship between K_C versus daily
857 PAR ($\text{mol quanta m}^{-2} \text{d}^{-1}$) and NPQ Proxy (dimensionless). The linear equations are
858 the results from Type II regression.

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862 **Fig. 10 (a)** Comparisons of measured Chla-specific ^{13}C -uptake rates against
863 estimated from FRRf- based ETRs obtained in 2007 and K_C model (Fig. 9(c)) and **(b)**
864 measured depth-integrated primary productivities against FRRf-based estimated for
865 all cruises across 2007 – 2010.

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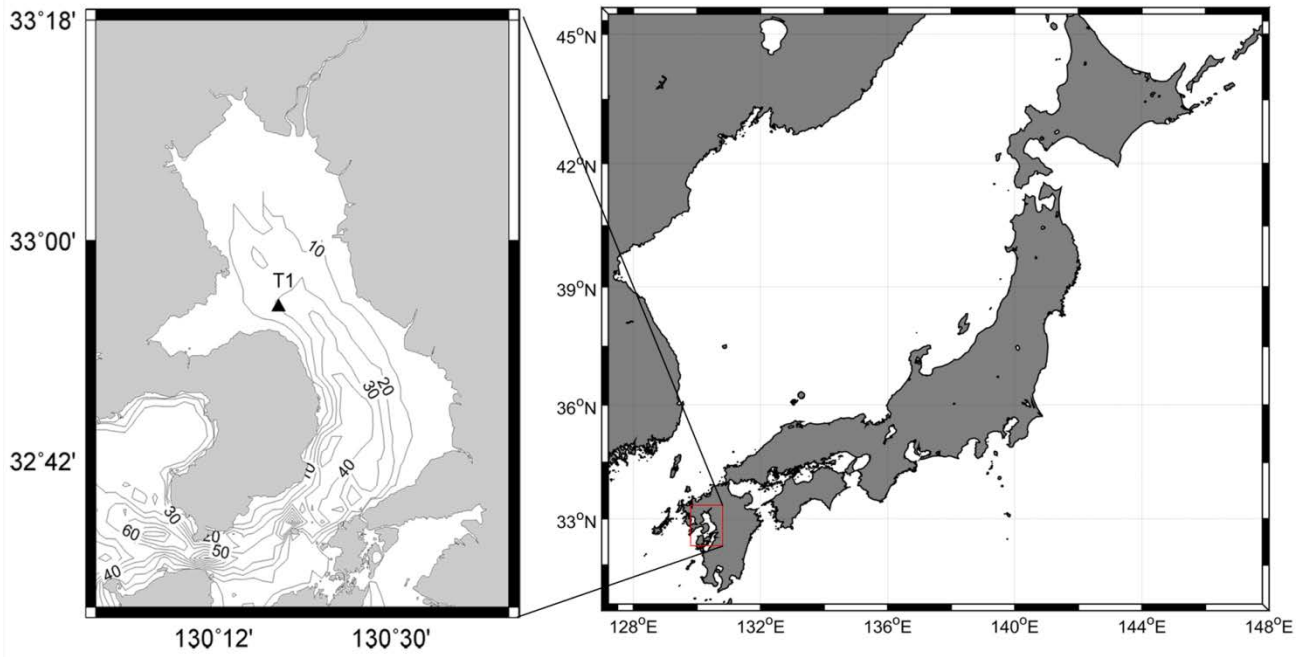


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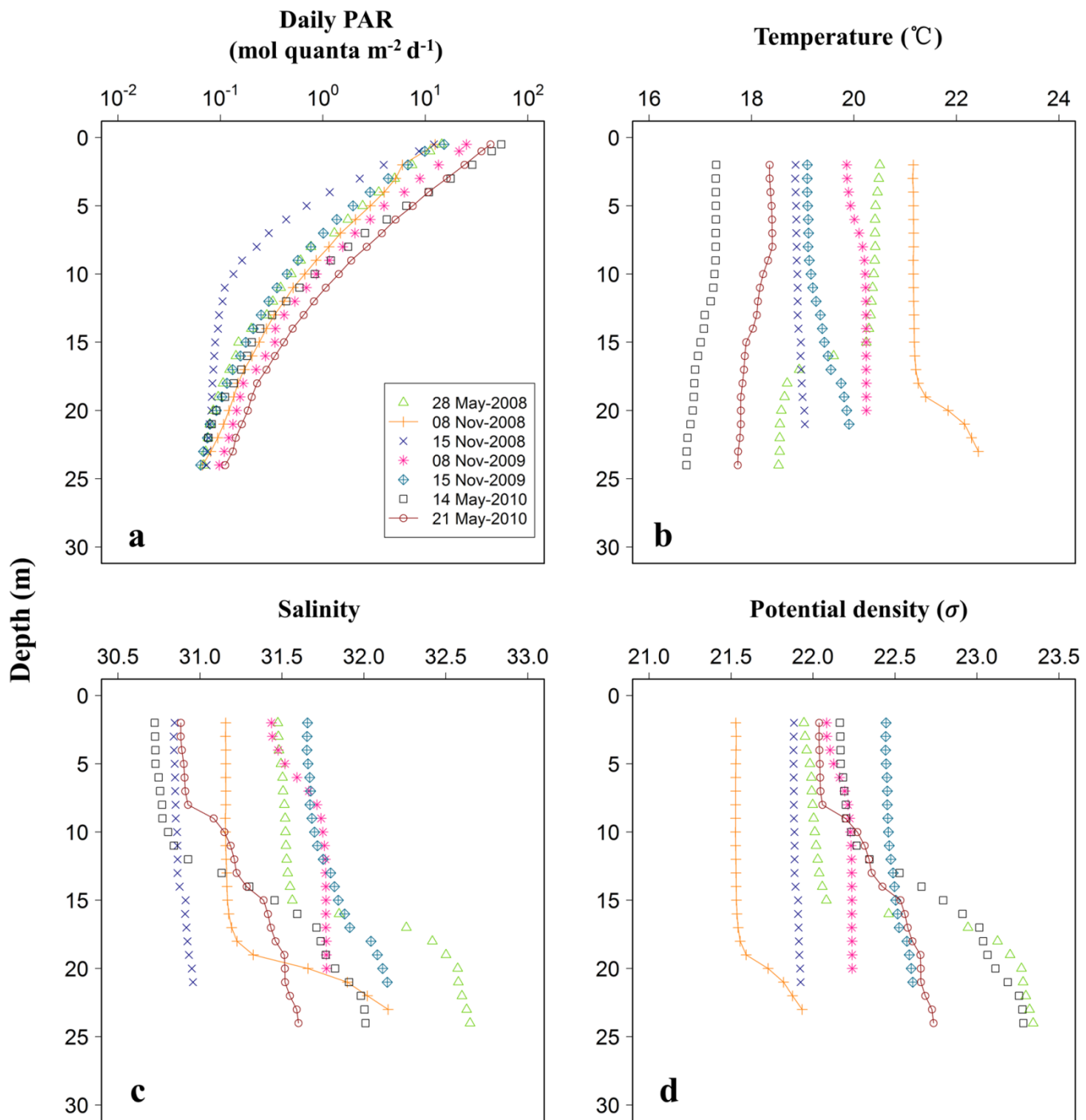


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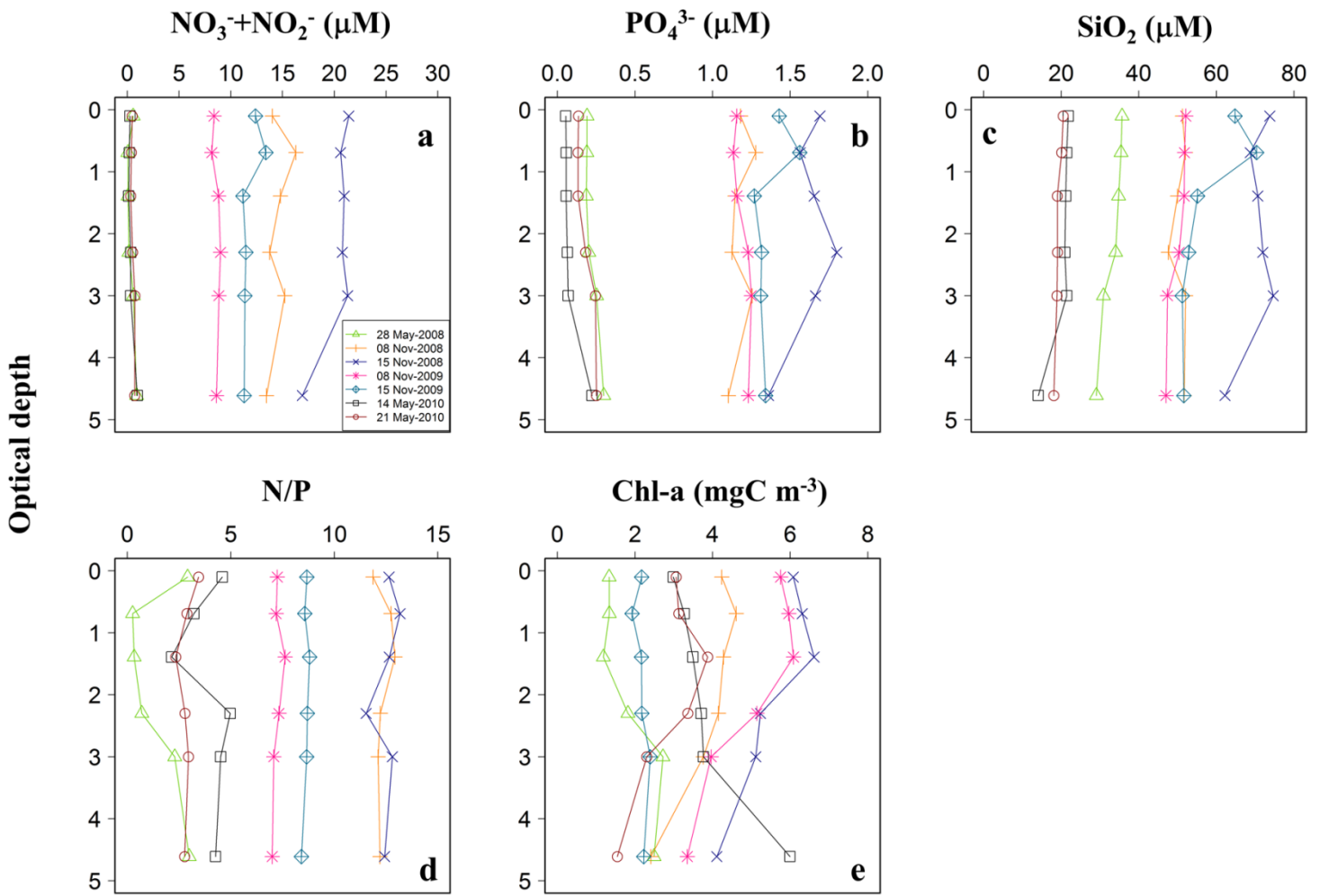


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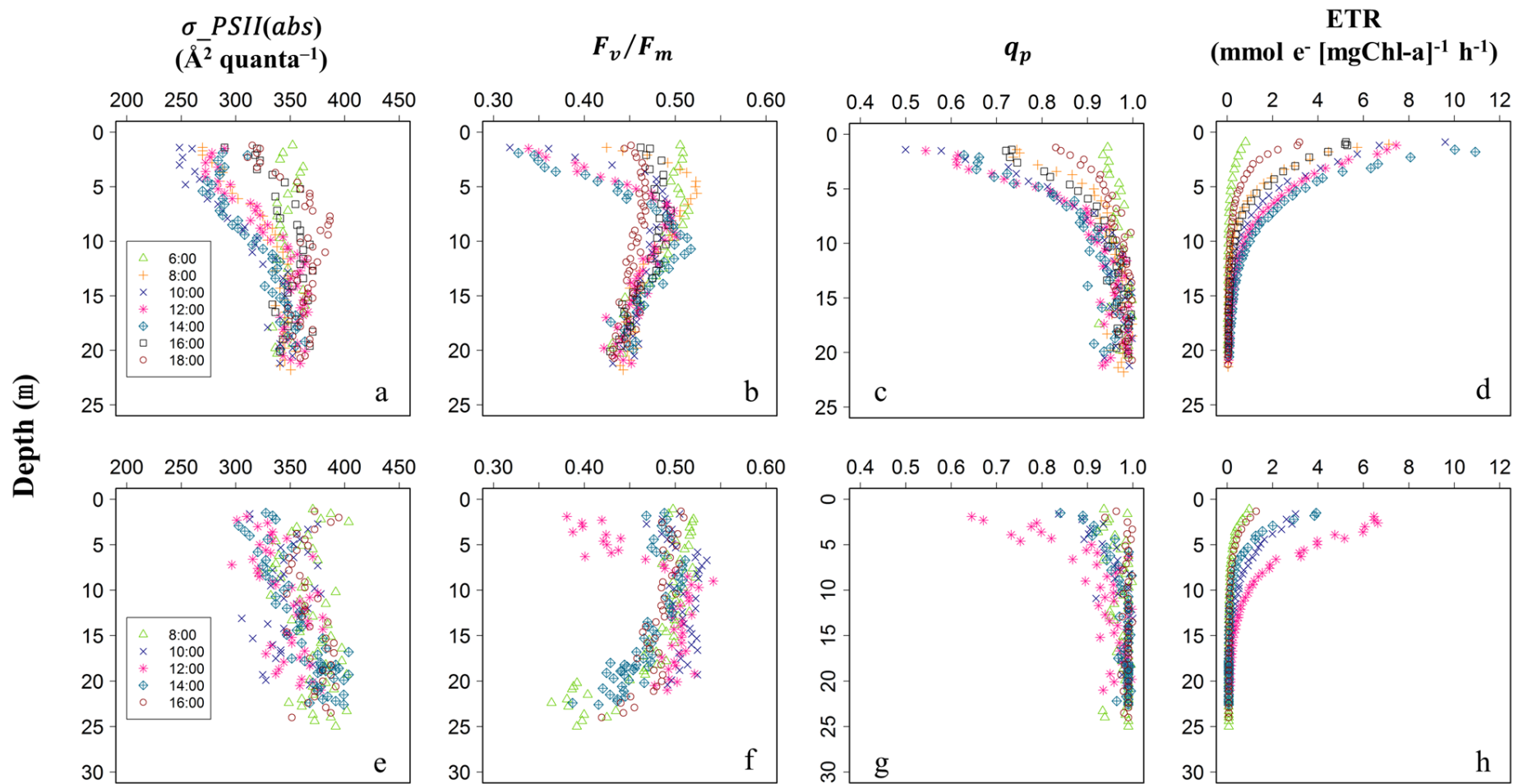


Fig. 4

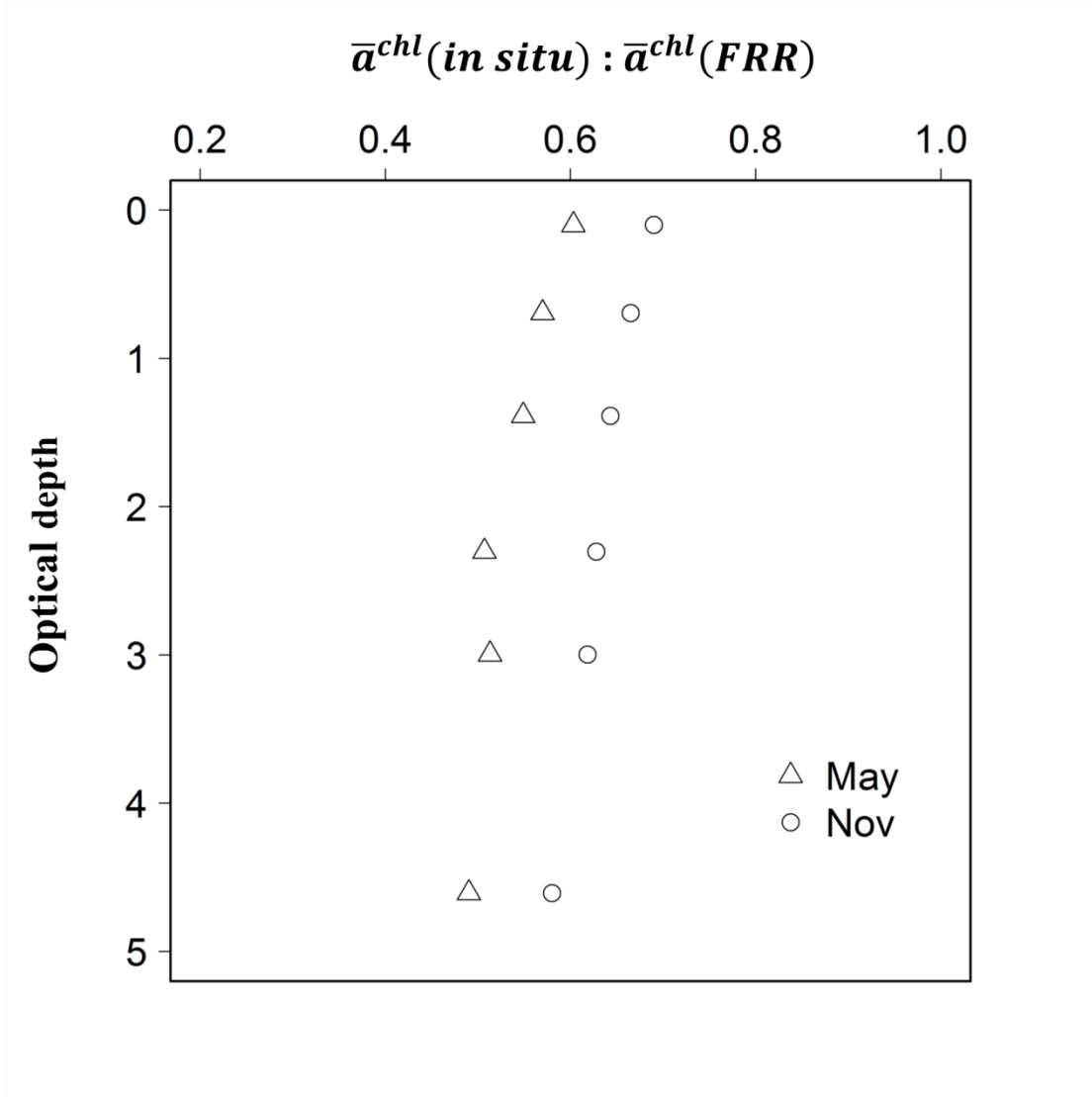


Fig. 5

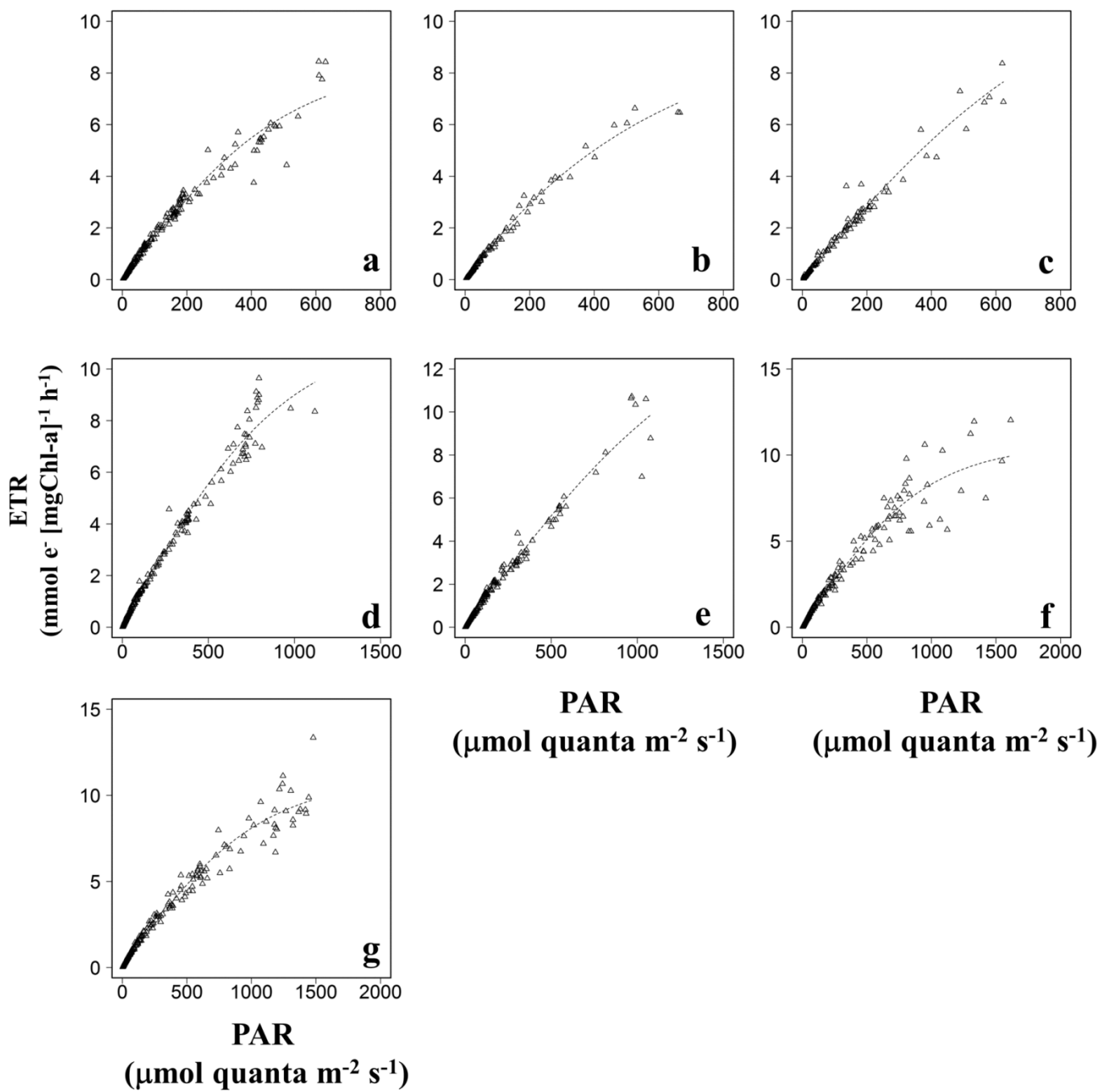


Fig. 6

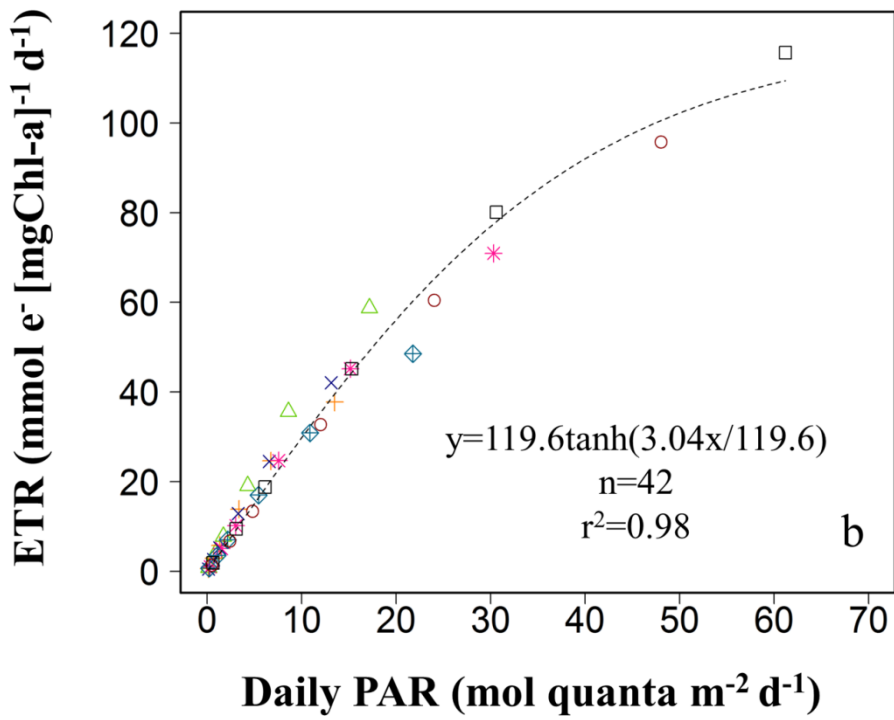
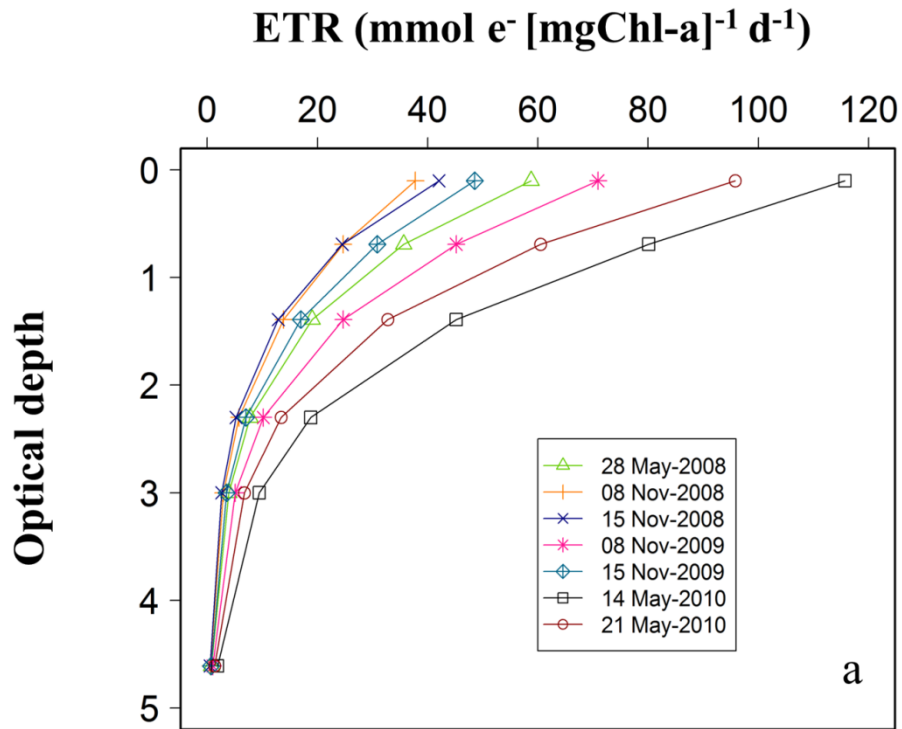


Fig. 7

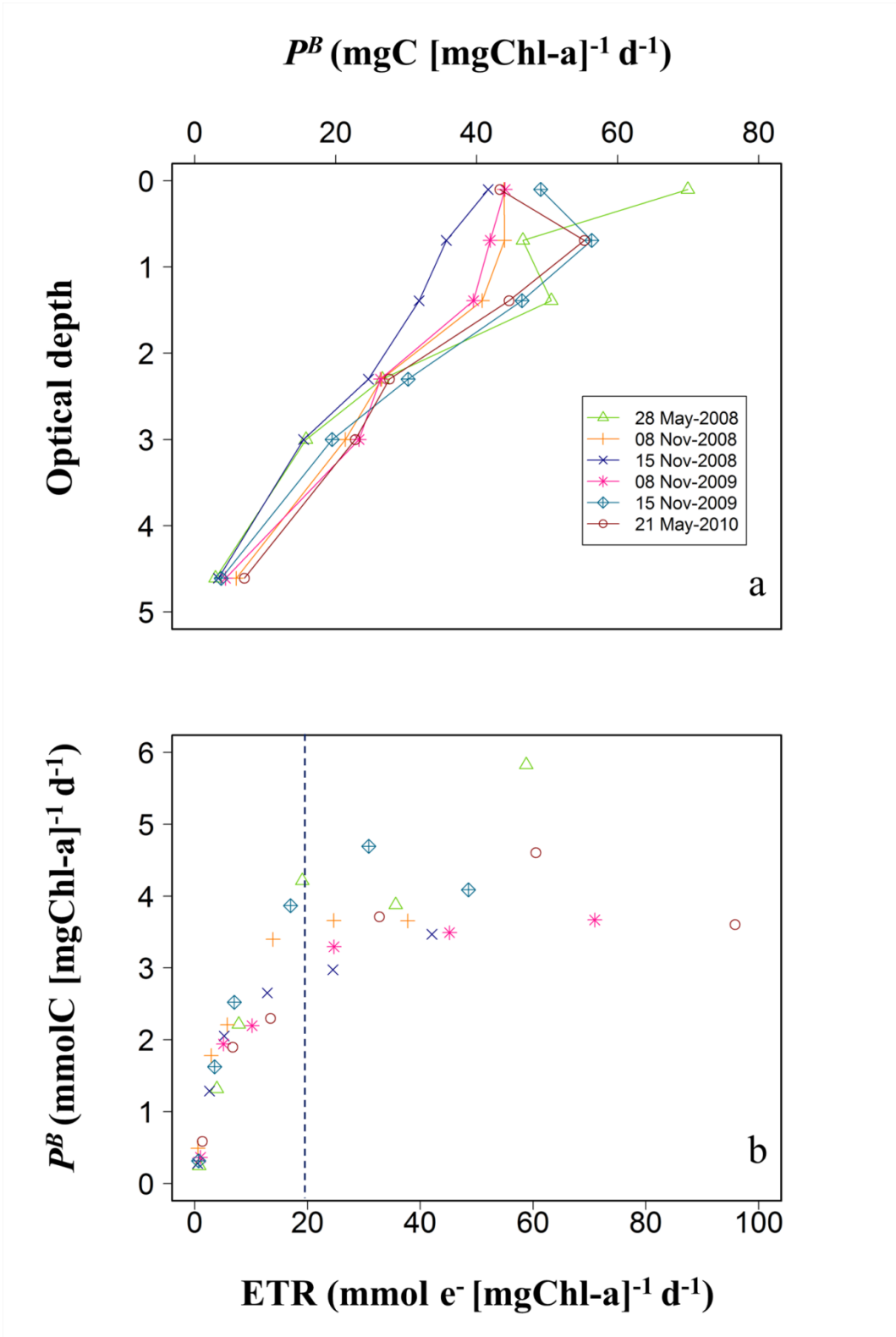


Fig. 8

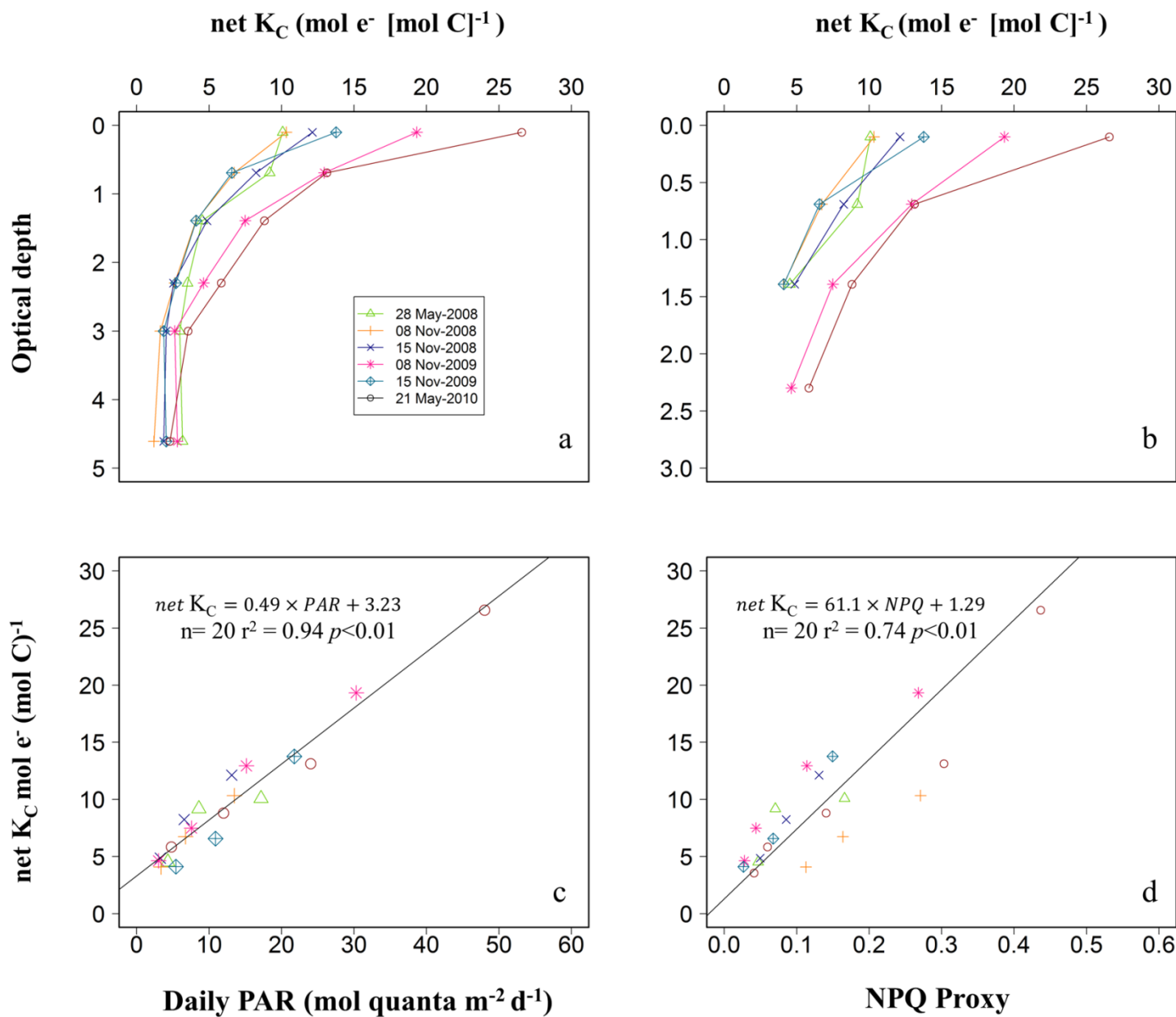


Fig. 9

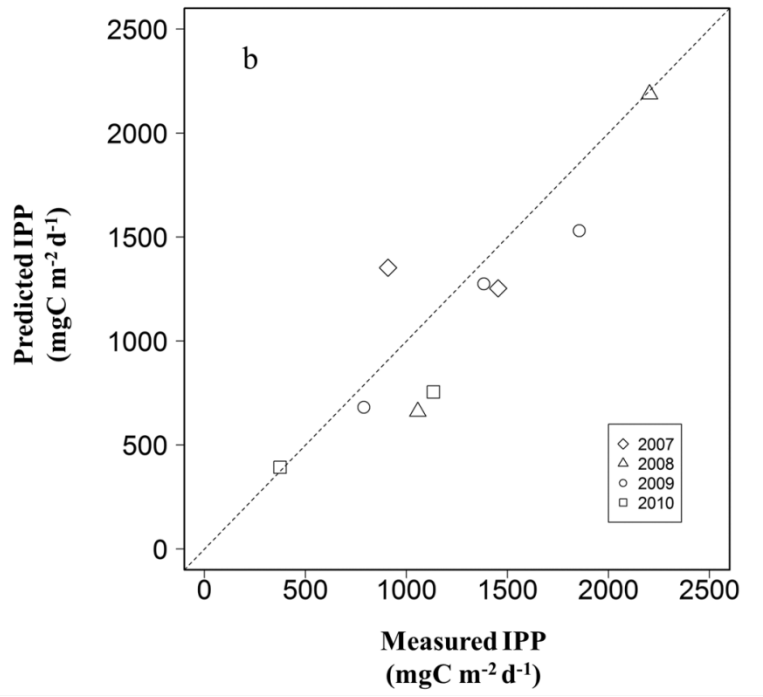
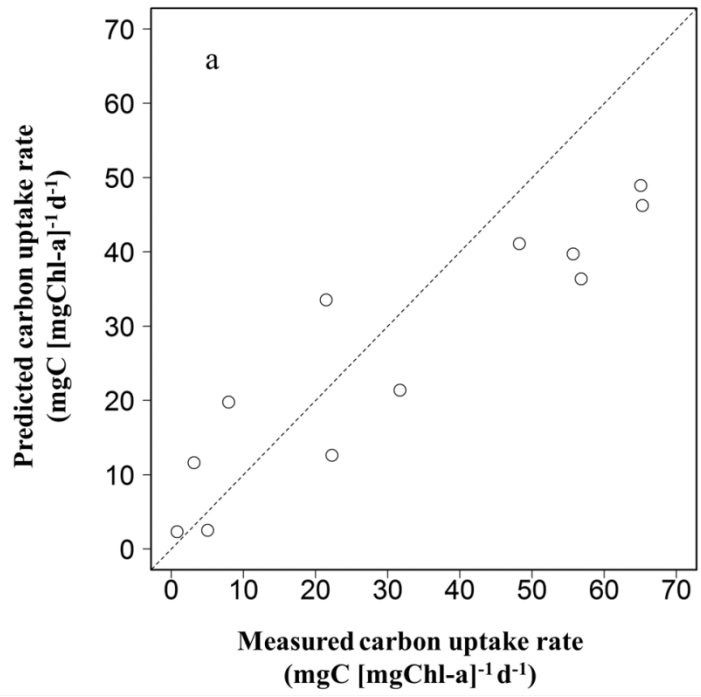


Fig. 10